

**Removal of *Microcystis aeruginosa* by UV-activated persulfate  
Performance and characteristics**

Wang, Zongping; Chen, Yiqun; Xie, Pengchao; Shang, Ran; Ma, Jun

**DOI**

[10.1016/j.cej.2016.04.125](https://doi.org/10.1016/j.cej.2016.04.125)

**Publication date**

2016

**Document Version**

Accepted author manuscript

**Published in**

Chemical Engineering Journal

**Citation (APA)**

Wang, Z., Chen, Y., Xie, P., Shang, R., & Ma, J. (2016). Removal of *Microcystis aeruginosa* by UV-activated persulfate: Performance and characteristics. *Chemical Engineering Journal*, 300, 245-253.  
<https://doi.org/10.1016/j.cej.2016.04.125>

**Important note**

To cite this publication, please use the final published version (if applicable).  
Please check the document version above.

**Copyright**

Other than for strictly personal use, it is not permitted to download, forward or distribute the text or part of it, without the consent of the author(s) and/or copyright holder(s), unless the work is under an open content license such as Creative Commons.

**Takedown policy**

Please contact us and provide details if you believe this document breaches copyrights.  
We will remove access to the work immediately and investigate your claim.

1 **Removal of *Microcystis aeruginosa* by UV-activated persulfate: Performance and**  
2 **characteristics**

3  
4 Zongping Wang<sup>a,b</sup>, Yiqun Chen<sup>a,b</sup>, Pengchao Xie<sup>a,b,\*</sup>, Ran Shang<sup>c,\*\*</sup>, and Jun Ma<sup>d</sup>

5 *<sup>a</sup> School of Environmental Science and Engineering, Huazhong University of Science*  
6 *and Technology (HUST), Wuhan 430074, China*

7 *<sup>b</sup> Key Laboratory of Water and Wastewater Treatment (HUST), MOHURD, Wuhan,*  
8 *430074, China*

9 *<sup>c</sup> Department of Sanitary Engineering, Faculty of Civil Engineering and Geosciences,*  
10 *Delft University of Technology, P.O. Box 5048, 2600 GA Delft, The Netherlands*

11 *<sup>d</sup> State Key Laboratory of Urban Water Resource and Environment, Harbin Institute*  
12 *of Technology, Harbin 150090, China*

13 **Abstract**

14 Cyanobacteria blooms in source waters have become a worldwide issue for drinking  
15 water production. UV-activated persulfate (UV/PS) technology was firstly applied to  
16 remove cultivated *Microcystis aeruginosa* in bench scale. The presence of persulfate  
17 significantly enhanced both cytolysis and algal organic matter mineralization  
18 compared with UV-C inactivation alone. Around 98.2% of algal cells were removed  
19 after UV/PS process treatment for 2 h at a dosage of PS being 1500 mg/L  
20 (approximately 6 mM). Both sulfate and hydroxyl radicals were proven to contribute

---

\* Corresponding authors.

E-mail address: [pengchao.x@gmail.com](mailto:pengchao.x@gmail.com) (P. Xie), [r.shang@tudelft.nl](mailto:r.shang@tudelft.nl) (R. Shang).

21 to the removal of algae and the loss of cell integrity. The cultivated *Microcystis*  
22 *aeruginosa* in death growth phase were found to be more vulnerable to UV/PS  
23 treatment than those growing in log phase, thus a significant lower dosage of PS is  
24 needed to achieve the desired removal efficiency. This study suggested a novel  
25 application of UV/PS process in the removal of algae in source waters due to the high  
26 degradation efficiency of both algal cells and their derived organic matter.

27

28 *Keywords: Microcystis aeruginosa; Ultraviolet; Persulfate; Cell integrity; Algal*  
29 *organic matter*

30

31

## 32 **1. Introduction**

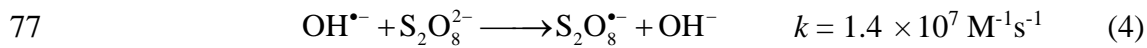
33 In recent decades, harmful algal blooms have frequently occurred in source waters  
34 such as lakes, rivers and reservoirs associated with eutrophication throughout the  
35 world [1-4]. *Cyanobacteria*, a prominent and ubiquitous issue, has attracted  
36 worldwide attention among the harmful algal blooms. Along with excessive algal cells,  
37 the algal organic matter (AOM) including extracellular organic matter (EOM) and  
38 intracellular organic matter (IOM) generated via metabolic excretion always cause  
39 serious water problems [5]. The AOM produced by some specific *Cyanobacteria*  
40 genera including *anabaena*, *microcystis*, *planktothrix*, etc. has been proven to  
41 comprise taste- and odor- substances, such as 2-methylisoborneol and geosmin [6],  
42 and a wide range of toxic intracellular metabolites being suggested to cause both acute  
43 and chronic effects on hepatocyte and central nervous system of aquatic organisms  
44 and biomagnify [7-10]. Moreover, algal cells and AOM have been reported to be  
45 important precursors of vast disinfection by-products (DBPs) including traditional  
46 trihalomethanes (THMs), haloacetic acids (HAAs) and emerging nitrogenous DBPs  
47 (N-DBPs) containing high genotoxicity and carcinogenicity [11-13].

48 However, the conventional drinking water treatment plant only shows limited  
49 removal efficiency on algal cells due to electrostatic repulsion, surface hydrophilicity  
50 and steric effects [14-16]. The residual cells after sedimentation could adhere to the  
51 filter material surface subsequently causing filter clogging, penetrate into the water  
52 supply pipe network, and finally impair the drinking water quality [17]. It should be

53 noted that the dissolved AOM generated via metabolic excretion are even more  
54 different than algal cells to be removed by using traditional  
55 coagulation-sedimentation-filtration process [18], which may adversely affect  
56 conventional water production via inhibition of coagulation [19, 20].

57 Activated persulfate (PS) oxidation has been studied as an alternative conventional  
58 advanced oxidation process (AOP) in water treatment [21-23]. The AOP using PS is  
59 mainly achieved by the formation of reactive sulfate radical ( $\text{SO}_4^{\bullet-}$ ,  $E^0=2.65\text{--}3.1\text{ V}$ )  
60 through the decomposition of PS by heat, transit metals, light, microwave or  
61 ultrasound [24-26]. Similar to hydroxyl radical ( $\text{HO}^{\bullet}$ ,  $E^0=1.8\text{--}2.7\text{ V}$ ), electron-transfer  
62 is expected to be a vital reaction when  $\text{SO}_4^{\bullet-}$  is used to degrade organic pollutants [27].  
63 However,  $\text{SO}_4^{\bullet-}$  is more selective in comparison with hydroxyl radical in general, thus  
64 might be more effective in the degradation of some organic pollutants in the presence  
65 of radical scavengers [27]. Particularly, PS activated by zerovalent iron was recently  
66 used for disinfection of ballast water and achieved a result that the species of marine  
67 phytoplankton could be inactivated by such a process without generating harmful  
68 byproducts [28]. However, to our knowledge, activated PS technology has not been  
69 utilized for algae removal in source waters.

70 Among the known PS-activation processes, ultraviolet (UV) radiation at 254 nm  
71 can activate the PS to generate  $\text{SO}_4^{\bullet-}$  through Eq. 1 with a quantum yield of 0.7 mol  
72 E/s [29], and the other main chemical interactions subsequently take place were  
73 summarized in Eqs. 2 ~ 5 [30].



79 Furthermore, UV process has been applied to suppress algae growth in many cases  
 80 accompanied by chlorophyll bleaching and inhibition of metabolic activity [31-35].  
 81 Therefore, the UV radiation can be regarded as one of the feasible techniques to  
 82 activate the PS for drinking water treatment applications.

83 In the present study, the performance of UV-activated PS (UV/PS) process on algae  
 84 removal has been investigated using *Microcystis aeruginosa* (*M. aeruginosa*), a most  
 85 abundant and common occurring cyanobacteria specie [36]. Additionally, variations in  
 86 the characteristics of AOM during the UV/PS treatment were identified to further  
 87 recognize the plausible by-products within the oxidation process.

88

## 89 **2. Materials and methods**

### 90 **2.1. Materials**

91 The *M. aeruginosa* (No. FACHB-909) was purchased from the Institute of  
 92 Hydrobiology, Chinese Academy of Sciences, and laboratorial cultivated in a light  
 93 growth incubator (Guohua Electric Co., Ltd., China). The *M. aeruginosa* was  
 94 incubated under the specific growth conditions ( $25 \pm 1$  °C with a light-dark cycle of

95 12 h: 12 h) in BG-11 media [37]. The algae solutions in log phase were harvested and  
96 diluted using ultrapure water (18.5 M $\Omega$  cm) produced from a water purifier (PCDX-J,  
97 Pincheng Co. Ltd, China). Then a final cell density of  $1 \times 10^6$  cell/mL was achieved  
98 for the experiments, which mimics to the practical density in harmful algal blooms  
99 [38].

100 All chemicals used in the experiments were of analytical reagent grade at least.  
101 Sodium persulfate (Na<sub>2</sub>S<sub>2</sub>O<sub>8</sub>), sodium hydroxide, hydrochloric acid, methanol (MeOH)  
102 and tert-butyl-alcohol (TBA) were obtained from Sinopharm Chemical Reagent Co.,  
103 Ltd., China. Powdered Na<sub>2</sub>S<sub>2</sub>O<sub>8</sub> was added to the prepared *M. aeruginosa* solutions to  
104 achieve designated concentrations of PS each time. Solution pH was subsequently  
105 adjusted to be 7.0 which is around the pH of natural water by using HCl and NaOH at  
106 a concentration of 0.1 M.

107

## 108 **2.2. Experimental reactor and procedures**

109 Algae removal experiments were carried out in a cylindrical pyrex reactor (600 mL  
110 effective volume) with a low-pressure UV lamp (254 nm, 23W, GPH 436T5L/4,  
111 Philips Electronics Ltd., The Netherlands) (Fig. S1). An immersion well made of high  
112 purity quartz was placed inside the glass reactor. The UV lamp was fixed inside the  
113 immersion well. Cooling water was pumped through the thin annular zone of the  
114 immersion well to prevent overheating of the reaction solutions. In order to achieve a  
115 stabilized radiation intensity (measured to be 1.25 mW/cm<sup>2</sup> in average by the reported

116 method [39]), the lamp was always switched on for 15 min before being placed into  
117 the reactor. A magnetic stirring apparatus at a speed of 200 rpm was used to  
118 homogenize the solutions throughout the experiments. Samples were collected via the  
119 sampling port at specific time intervals. To evaluate the reaction mechanisms in the  
120 UV/PS system, MeOH and TBA were added as scavengers for hydroxyl and sulfate  
121 radicals. Each batch of experiment was carried out in triplicate. As the formed  $\text{SO}_4^{\cdot-}$   
122 could be scavenged by high concentrations of  $\text{S}_2\text{O}_8^{2-}$  and  $\text{Cl}^-$  from BG-11 media and  
123 HCl solution (section 3.2), the presence of anions including  $\text{NO}_3^-$  and  $\text{HCO}_3^-$  at  
124 concentrations as usual in natural waters was expected to play little role in the  
125 removal of algal cells (Fig. S2). Thus, the impacts of co-existing anions would not be  
126 further discussed in the following sections.

127

### 128 **2.3. Analytical methods**

129 The cell concentration of *M. aeruginosa* was measured using an UV-visible  
130 spectrophotometer (U-3100, Hitachi, Japan) at a wavelength of 681 nm since the  
131 optical density at 681 nm ( $\text{OD}_{681}$ ) is linearly correlated with counted cell number by  
132 microscope within the experimental range [40, 41]. Thus, the removal efficiency of  
133 algal cells ( $\rho$ , %) can be calculated using Eq. 6.

$$134 \quad \rho = \frac{(\text{OD}_{681_0} - \text{OD}_{681_t})}{\text{OD}_{681_0}} \times 100\% \quad (6)$$

135 where  $\text{OD}_{681_0}$  and  $\text{OD}_{681_t}$  were the optical density values at 0 min and  $t$  min.

136 Chlorophyll-a (Chl-a) was extracted using acetone solution and then measured



137 using the spectrophotometer at wavelengths of 663 nm, 645 nm and 630 nm according  
138 to the reported method [42].

139 The algal cell integrity before and after treatment was monitored by a flow  
140 cytometer (Accuri C6, BD Biosciences, USA) equipped with an argon laser emitting  
141 (wavelength fixed at 488 nm) for fluorescence measurement. Algal cells were stained  
142 using SYTOX green nucleic acid stain (Invitrogen, Life Technologies, USA) [43].  
143 Fluorescent filters and detectors were equipped to collect green fluorescence in  
144 channel FL1 (530 nm) and red fluorescence in channel FL3 (630 nm), which  
145 represents damaged and integrated cells, respectively, since SYTOX could penetrate  
146 damaged cells and stain the nucleic acid to emit green fluorescence.

147 Extracellular AOM was extracted from reaction solution by centrifuging the cell  
148 suspension at 8000 rpm for 15 min and subsequently filtering the supernatant through  
149 0.45  $\mu\text{m}$  cellulose acetate membranes [5]. The extracellular AOM was characterized  
150 on fluorescence spectroscopy, UV-Vis spectrum scan, and total organic carbon (TOC).  
151 A fluorescence spectrophotometer (F-4600, Hitachi, Japan) was used to measure the  
152 fluorescence excitation–emission matrix (EEM) spectroscopy of AOM. Excitation  
153 wavelengths (Ex) were scanned from 200 to 450 nm with 5 nm intervals and emission  
154 wavelengths (Em) from 280 to 550 nm with 2 nm intervals. The scanning speed was  
155 set at 1200 nm/min. Background signals were minimized by subtracting the signals of  
156 the blank (i.e. ultrapure water). The EEM data were analyzed by MATLAB 2010b  
157 (The MathWorks, Inc., USA). The UV-Vis spectrum scan of AOM ranging from 200

158 to 700 nm was measured using the U-3100 spectrophotometer. The concentration of  
159 dissolved organic carbon (DOC) in the prepared sample was measured using a  
160 TOC/TN analyzer (C/N 2100, Analytic Jena, Germany).

161 The persulfate concentration was measured by universal iodometric titration  
162 method [44]. The procedure was conducted by mixing 2 mL samples and 10 mL KI  
163 stock solution (10% mass fraction) in 40 mL ultrapure water. Then the resulting  
164 solutions were equilibrated for 3 h in dark, and subsequently titrated by calibrated  
165 Na<sub>2</sub>S<sub>2</sub>O<sub>3</sub> solution (0.1 M). A turbidimeter (Ruixin WGZ-2, China) was used to  
166 measure the turbidity of algal solution before and after treatment. The *M. aeruginosa*  
167 cells in suspension for morphology observation were firstly centrifuged at 6000 rpm  
168 to collect the precipitate, and then dried by a vacuum freeze dryer (model FD-1A-50,  
169 Shanghai Boyikang Instrument Co., Ltd., China). Thereafter, the dried algae samples  
170 were sputter coated with gold by the sputter coater, and then photographed using a  
171 scanning electron microscopy (SEM) (Sirion 200, FEI, USA) at 10 kV.

172

### 173 **3. Results and discussion**

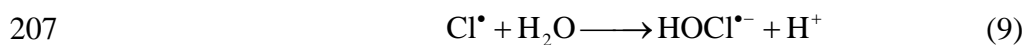
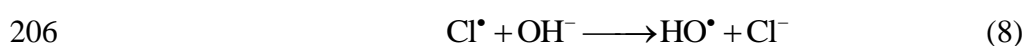
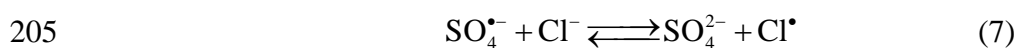
#### 174 ***3.1. Comparison of UV, PS and UV/PS processes on algae removal***

175 Fig. 1 shows the removal of *M. aeruginosa* based on the changes of OD<sub>681</sub> (Fig. 1a)  
176 and Chl-a (Fig. 1b) by using three different oxidation processes, UV irradiation, PS  
177 oxidation and combined UV/PS oxidation, at 25 °C with the initial reaction pH being  
178 7.0. Negligible decrease of OD<sub>681</sub> and Chl-a was observed using PS oxidation alone at

179 a concentration of 1500 mg/L (about 6 mM) over a period of 120 min, indicating little  
180 to no oxidation of algal cells by PS directly. In contrast, UV irradiation could degrade  
181 algal cells to some extent with OD<sub>681</sub> and Chl-a being removed by 21.8% and 87.9%,  
182 respectively, after 120 min of radiation. The results corroborate that UV at 254 nm can  
183 effectively destroy the photosynthesis system of *M. aeruginosa*, but hardly cause  
184 large-scale cell cytolysis, which was also reported in other publications [32, 45].  
185 UV/PS process showed the highest levels of removal efficiency for both OD<sub>681</sub>  
186 (98.2%) and Chl-a (99.8%) among the three oxidation processes.

187 As PS oxidation and UV irradiation had limited degradation efficiency on algal  
188 cells, it could be expected that the reactive radicals, i.e. SO<sub>4</sub><sup>•-</sup> and HO<sup>•</sup>, produced in  
189 UV/PS system were responsible for the observed degradation of algal cells. The  
190 conclusion was further proved by the results shown in Fig. S3 that the presence of  
191 MeOH and TBA inhibited the removal of OD<sub>681</sub> significantly when using UV/PS  
192 process. MeOH is usually considered as an effective quencher for both the SO<sub>4</sub><sup>•-</sup> and  
193 HO<sup>•</sup> [46]. However, TBA is an effective quencher for HO<sup>•</sup>, but not for SO<sub>4</sub><sup>•-</sup> [29].  
194 Thus, they could be used to distinguish SO<sub>4</sub><sup>•-</sup> and HO<sup>•</sup> based on the difference of the  
195 degradation rate of target compounds. Fig. S3 shows that co-existence of either  
196 MeOH or TBA inhibited the removal of OD<sub>681</sub>, while the inhibiting ability of MeOH  
197 is stronger than that of TBA. It proved that both SO<sub>4</sub><sup>•-</sup> and HO<sup>•</sup> contributed to the  
198 oxidation of algal cells in the experiment. It is known that activation of PS by UV  
199 only generate SO<sub>4</sub><sup>•-</sup> (Eq. 1) without the production of HO<sup>•</sup> [29]. However, HO<sup>•</sup> was

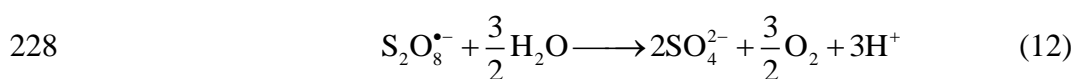
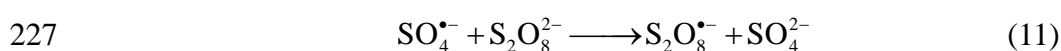
200 proposed to be generated when  $\text{SO}_4^{\bullet-}$  reacts with  $\text{OH}^-$  at a rate constant of  $6.5 \times 10^7$   
 201  $\text{M}^{-1} \text{S}^{-1}$  through Eq. 4 [30], or with water at a rate constant of  $8.3 \text{ M}^{-1} \text{S}^{-1}$  through Eq.  
 202 3 [47]. Besides, the presence of chloride could convert  $\text{SO}_4^{\bullet-}$  to  $\text{HO}^\bullet$  through reactions  
 203 shown in Eqs. 7 ~ 10 [48], especially when the reaction pH was higher than 5.0,  
 204 which make sure the presence of  $\text{HO}^\bullet$  in the experiment [49].



### 209 **3.2. Effect of initial PS dose on algae removal**

210 The effect of initial PS dose ranging from 0 mg/L to 1500 mg/L on the removal of *M.*  
 211 *aeruginosa* by focusing on the changes of  $\text{OD}_{681}$  and Chl-a was investigated (Fig. 2).  
 212 The removal rates of  $\text{OD}_{681}$  and Chl-a increased with the increase of initial PS dosage.  
 213 When the dosage of PS increased from 0 mg/L to 1500 mg/L (about 6 mM), the  
 214 removal rates of  $\text{OD}_{681}$  and Chl-a increased from 21.8% and 87.9% to 98.2% and  
 215 99.8%, respectively, after treatment for 120 min. It should be noted that, the rapid  
 216 decrease of Chl-a was observed in the initial 60 min treatment (including the only UV  
 217 irradiation case), but not for the removal of  $\text{OD}_{681}$ . The result suggested that the  
 218 photosynthetic system of *M. aeruginosa* was damaged immediately through  
 219 synergetic irradiation and oxidation in UV/PS system, possibly due to the rapid  
 220 destruction in gene expression of both *psbA* (for D1) and *cpc* (for phycocyanin) [50].

221 The variation of pHs over time under different initial PS doses was displayed in Fig.  
 222 S4. Negligible change of pHs was observed under UV irradiation alone over a period  
 223 of 120 min. However, solution pHs gradually reduced in UV/PS system during the  
 224 oxidation, which was enhanced with the increase of PS doses. The results can be  
 225 explained by the production of sulfate acid through  $S_2O_8^{\bullet-}$  oxidation of water (Eqs. 11  
 226 and 12) [30, 51]:



229 Fig. S5 shows that the dissolved oxygen (DO) in the experimental solution  
 230 significantly increased in the initial 30 min when using UV/PS process to treat algae,  
 231 supporting the speculation on oxidation of water by  $S_2O_8^{\bullet-}$ . Then the gradually  
 232 reduction of DO in the subsequent 90 min might due to the release into air, reacting  
 233 with  $HO^{\bullet}$  adducts to produce peroxy transients [52], and being converted to  
 234 oxygen-centered radicals which could be consumed by algae and its derived organic  
 235 matter [53].

236 If the dosed persulfate was completely decomposed, formed sulfate at  
 237 concentrations of 800 mg/L and 1200 mg/L would be achieved when the dosages of  
 238 persulfate were 1000 mg/L and 1500 mg/L, respectively. A sulfate concentration that  
 239 is higher than 600 mg/L in drinking water might cause taste alteration and diarrhea  
 240 [54]. However, it should be noted that the concentrations of *M. aeruginosa* in most  
 241 source waters were always lower than  $10^6$  cells/mL. When the initial cell density was

242 reduced to  $2.3 \times 10^5$  cells/mL, the residual OD<sub>681</sub> and turbidity decreased by 57.9%  
243 and 62.3%, respectively, after UV/PS process treatment for 120 min with an initial PS  
244 dosage of 200 mg/L (Fig. S6).

245

### 246 ***3.3. Effect of growth phase on algae removal***

247 The growth of *M. aeruginosa* can be modeled with four different phases: lag phase,  
248 log phase, stationary phase, and death phase. The removal of algae in their log phase  
249 and death phase were investigated due to the large variation of cell population and  
250 dissolved organic substances [55]. The effect of growth phase on the removal of *M.*  
251 *aeruginosa* was studied under different PS doses (Fig. 3). The removal of OD<sub>681</sub> in  
252 death phase kept high efficiencies (>92.8%) for all the initial concentrations of PS  
253 ranging from 0 mg/L to 1000 mg/L over an oxidation period of 120 min. However, for  
254 the cells in log phase, the removal of OD<sub>681</sub> only increased from 21.8% to 83.7% with  
255 the initial dosage of PS increasing from 0 mg/L to 1500 mg/L after reaction for 120  
256 min, which was much lower than the removal efficiencies for the algal cells in death  
257 phase. Similarly, the removal efficiency of Chl-a ranging from 90.7% to 98.5% in  
258 death phase by using UV/PS was also higher than that ranging from 87.9% to 97.6%  
259 in log phase.

260 The easier removal of *M. aeruginosa* in death phase than that in log phase by using  
261 UV/PS process was supposed to be connected to the disparity of biological activity,  
262 EOM component, and morphologies of algal cells in different growth phases [56].

263 Besides, the autolysis of cells largely occurred in death phase [19], which might also  
264 contribute to the high removal efficiency.

265

#### 266 **3.4. Mineralization of AOM**

267 Fig. 4 shows the residual TOC and extracellular DOC of *M. aeruginosa* suspension at  
268 different PS doses. Negligible changes of TOC and extracellular DOC were observed  
269 after single UV irradiation over a period of 120 min, indicating that limited  
270 UV-induced mineralization of AOM occurred, which was in accordance with the  
271 previous report [57]. On the other hand, the results reveal that the UV irradiation only  
272 had limited damage on the structure of algal cell and did not cause large-scale  
273 cytolysis [58], since an increase in DOC concentration would be found due to the  
274 release of IOM from damaged algal cells [25, 48].

275 With the dosage of PS increasing from 0 mg/L to 1500 mg/L (about 6 mM), the  
276 TOC and extracellular DOC were decreased from 12.50 and 7.59 mg/L to 1.63 and  
277 1.61 mg/L, respectively, over a reaction period of 120 min. The significant removal of  
278 extracellular DOC (decreased by 79.6% at 1500 mg/L PS dosage) indicates that  
279 UV/PS process can effectively mineralize AOM in the reaction solution. Moreover,  
280 the result of higher reduced concentration of TOC than extracellular DOC in the  
281 presence of PS can be explained by the fact that released dissolved IOM could make  
282 up parts of mineralized DOC, reflecting that UV/PS oxidation can induce destruction  
283 of algal cell structures. This speculation was supported by the result that residual TOC

284 (1.63 mg/L) and extracellular DOC (1.61 mg/L) were extremely approximate at an  
285 initial PS dosage of 1500 mg/L.

286 The UV-Vis spectra of *M. aeruginosa* solution before and after UV/PS treatment  
287 were also studied (insert graph in Fig. 4). The absorbance of solution ranging from  
288 200 to 700 nm was significantly decreased after UV/PS treatment, which was in  
289 accordance with the changes of TOC and DOC. It is noteworthy that, the peak at 681  
290 nm was disappeared after treatment, suggesting the effective removal of algal cells in  
291 UV/PS system.

292

### 293 ***3.5. Evolution in fluorescence EEM spectra of extracellular AOM***

294 The effects of UV/PS process treatment on fluorescence EEM spectra of extracellular  
295 AOM are shown in Fig. 5 and Table S1. There were four fluorescence peaks at Ex/Em  
296 wavelengths of 270/442 nm (peak A), 350/432 nm (peak C), 230/330 nm (peak T<sub>2</sub>),  
297 and 280/330 nm (peak T<sub>1</sub>) in the control sample (Fig. 5a), which represented  
298 fulvic-like, humic-like and protein-like substances, and dissolved microbial  
299 metabolites, respectively [59, 60]. The strong signal at peak T<sub>1</sub> suggested high  
300 concentration of protein-like substances contained in EOM, which was similar to  
301 other reports [61, 62].

302 However, after UV/PS process treatment for 5 min, peaks A and C were observed to  
303 be significantly enhanced, accompanied with the disappearance of peaks T<sub>1</sub> and T<sub>2</sub> in  
304 contrast (Fig. 5b). This can be explained that UV/PS process was supposed to cause



305 cell secretion and decomposition which was expected to produce fulvic-like and  
306 humic-like substances [63, 64].

307 When the treatment time was over 30 min, all the four peaks were disappeared  
308 (Figs. 5c and 5d), whose possible reasons were proposed as below: 1) UV/PS process  
309 could further oxidize and even mineralize the formed humic-like and fulvic-like  
310 substances (Fig. 4); and 2) the large-scale cell cytolysis and metabolites release  
311 mainly occurred in the initial reaction period under UV/PS oxidation, which was  
312 similar to the ozonation of *cyanobacteria* [65].

313 Additionally, comparative experiments of single UV irradiation, PS alone and  
314 UV/PS process treatment on the changes of the fluorescence EEM spectra of  
315 extracellular AOM were conducted (Fig. S7). After treatment for 30 min, UV  
316 irradiation caused significant decrease of peak T<sub>1</sub>, but slight enhancement of peak C.  
317 While the treatment of algae by PS alone had negligible impact on the fluorescence  
318 EEM spectra of the extracellular AOM. In comparison with UV irradiation or PS  
319 oxidation alone, UV/PS treatment caused the most significant decline of all the peaks  
320 on the fluorescence EEM spectra. Thus, it is reasonable to conclude that the formed  
321 reactive species, such as hydroxyl and sulfate radicals, in UV/PS process contribute to  
322 the degradation of organic matter.

323

### 324 ***3.6. Cell integrity and cellular morphology***

325 Fig. 6 shows the impacts of PS oxidation, UV irradiation, and UV/PS treatment on

326 cell integrity by using a flow cytometer to measure changes in cell permeability and  
327 chlorophyll auto-fluorescence. P1 and P2 regions represent damaged and integrated  
328 cell population, respectively. In Fig. 6a, the proportion of live cells was 96.1%,  
329 reflecting that the vast majority of *M. aeruginosa* cells without any treatment were  
330 intact. Compared with the control sample, the damaged cells shown in P1 region was  
331 found to be only slightly increased by 1.8% after 2 days reaction with PS alone (Fig.  
332 6b), indicating that individual PS oxidation played little to no roles in destruction of  
333 *M. aeruginosa* cells . In contrast, UV was observed to be more effective in decreasing  
334 the live algal cells as shown in regional P2 of Fig. 6c that the amount of live algal  
335 cells was reduced. However, the amount of algal cells shown in regional P1 was also  
336 decreased. Since the green fluorescence in channel FL1 (Regional P1) was from the  
337 stained nucleic acid by SYTOX [43], the results can be explained by the fact that UV  
338 irradiation could damage nucleic acid [66, 67]. Among the four treatment processes,  
339 UV/PS process had the strongest ability to damage algal cells (Fig. 6d). After  
340 treatment for 30 min, the integrated cell population was significantly reduced,  
341 indicating that the reactive substances in UV/PS system, such as  $\text{SO}_4^{\bullet-}$  and  $\text{HO}^{\bullet}$ , can  
342 strongly damage the cell integrity. No signals were found in regional P1 suggested  
343 that UV/PS could further oxidize nucleic acid effectively due to the synergetic effect  
344 of UV irradiation and the generated reactive radicals.

345 The surface morphologies of *M. aeruginosa* cells before and after UV/PS treatment  
346 were observed using the SEM. Compared to the algal cells without any treatments

347 (Fig. 7a), the surface morphologies of cells were deformed and the cellular structure  
348 was completely lost after the UV/PS treatment (Fig. 7b). Some apparent materials  
349 were found to be released from the cells, which were probable to be amino acids and  
350 carbohydrates [28].

351

### 352 **3.7. Persulfate decomposition**

353 Fig. 8a shows the persulfate decay under different initial PS doses ranging from 500  
354 to 1500 mg/L. In each case, residual persulfate was less than 2% over a reaction  
355 period of 120 min. The decomposition of persulfate followed the first-order kinetics  
356 model (Fig. 8b), suggesting that the presence of algal cells and AOM had little impact  
357 on the self-decomposition of persulfate, which was supported by the insignificant  
358 oxidation potential of PS on algal cells and AOM compared to UV/PS process (Figs. 1  
359 and S7). The added PS was almost completely decomposed ( $\geq 99.9\%$ ) during the  
360 reaction when the initial concentration of PS was less than 1000 mg/L. Even at initial  
361 dosages of 1250 mg/L and 1500 mg/L, the average residual PS concentrations were  
362 only 2.14 mg/L (about 0.2%) and 23.6 mg/L (about 1.6%) after treatment for 120 min,  
363 respectively. Thus, it is concluded that application of UV/PS process in the removal of  
364 algae in natural water had little potential to cause undesired risk from the residual PS.

365

## 366 **4. Conclusions**

367 UV/PS process is demonstrated to be an effective technology for the removal of *M.*

368 *aeruginosa* in laboratory. The formed reactive species including  $\text{SO}_4^{\bullet-}$  and  $\text{HO}^\bullet$  were  
369 proven to be the most important reasons for the removal of algal cells. Increasing PS  
370 dosage could improve the removal of algal cells and the mineralization of AOM. With  
371 the addition of 1500 mg/L (about 6 mM) PS, removal rates up to 98.2% and 99.8% of  
372 algal cells and Chl-a, respectively, were achieved in UV/PS system over a reaction  
373 period of 120 min. *M. aeruginosa* in death phase was easier to be removed by using  
374 UV/PS treatment than that in log phase. Although the UV/PS process significantly  
375 damaged the structures of algal cell and caused the release of IOM, these derived  
376 organic compounds along with the dissolved EOM could be further mineralized in the  
377 subsequent reaction period. The fast decomposition of persulfate might insure the  
378 safety of using UV/PS to remove algal cells to some extent.

379

### 380 **Acknowledgments**

381 This work was supported by the National Natural Science Foundation of China  
382 (Grant Nos. 51578259). We also thank the Analytical and Testing Center of Huazhong  
383 University of Science and Technology for related analysis.

384

### 385 **Appendix A. Supplementary data**

386 Supplementary data (Figs. S1 to S7 and Tab. S1) associated with this article can be  
387 found, in the online version.

388

389 **References**

- 390 [1] B.E. Lapointe, L.W. Herren, D.D. Debortoli, M.A. Vogel, Evidence of  
391 sewage-driven eutrophication and harmful algal blooms in Florida's Indian River  
392 Lagoon, *Harmful Algae* 43 (2015) 82-102.
- 393 [2] J.D. Isaacs, W.K. Strangman, A.E. Barbera, M.A. Mallin, M.R. McIver, J.L.C.  
394 Wright, Microcystins and two new micropeptin cyanopeptides produced by  
395 unprecedented *Microcystis aeruginosa* blooms in North Carolina's Cape Fear River,  
396 *Harmful Algae* 31 (2014) 82-86.
- 397 [3] M.A. Burford, S.A. Johnson, A.J. Cook, T.V. Packer, B.M. Taylor, E.R. Townsley,  
398 Correlations between watershed and reservoir characteristics, and algal blooms in  
399 subtropical reservoirs, *Water Research* 41 (2007) 4105-4114.
- 400 [4] V.H. Smith, Eutrophication of freshwater and coastal marine ecosystems: a global  
401 problem, *Environmental Science and Pollution Research* 10 (2003) 126.
- 402 [5] R.K. Henderson, A. Baker, S.A. Parsons, B. Jefferson, Characterisation of  
403 algogenic organic matter extracted from cyanobacteria, green algae and diatoms,  
404 *Water Research* 42 (2008) 3435-3445.
- 405 [6] J.O. Jo, S.D. Kim, H.J. Lee, Y.S. Mok, Decomposition of taste-and-odor  
406 compounds produced by cyanobacteria algae using atmospheric pressure plasma  
407 created inside a porous hydrophobic ceramic tube, *Chemical Engineering Journal* 247  
408 (2014) 291-301.
- 409 [7] J.P. Berry, P.D. Gibbs, M.C. Schmale, M.L. Saker, Toxicity of cylindrospermopsin,

410 and other apparent metabolites from *Cylindrospermopsis raciborskii* and  
411 *Aphanizomenon ovalisporum*, to the zebrafish (*Danio rerio*) embryo, *Toxicon* :  
412 official journal of the International Society on Toxinology 53 (2009) 289-299.

413 [8] R. Bhattacharya, K. Sugendran, R.S. Dangi, P.V.L. Rao, Toxicity evaluation of  
414 freshwater cyanobacterium *Microcystis aeruginosa* PCC 7806: II Nephrotoxicity in  
415 rats, *Biomedical and Environmental Sciences* 10 (1997) 93-101.

416 [9] C. MacKintosh, K.A. Beattie, S. Klumpp, P. Cohen, G.A. Codd, Cyanobacterial  
417 microcystin-LR is a potent and specific inhibitor of protein phosphatases 1 and 2A  
418 from both mammals and higher plants, *FEBS letters* 264 (1990) 187-192.

419 [10] M.J. Bakke, T.E. Horsberg, Effects of algal-produced neurotoxins on metabolic  
420 activity in telencephalon, optic tectum and cerebellum of Atlantic salmon (*Salmo*  
421 *salar*), *Aquatic toxicology* 85 (2007) 96-103.

422 [11] J. Fang, J. Ma, X. Yang, C. Shang, Formation of carbonaceous and nitrogenous  
423 disinfection by-products from the chlorination of *Microcystis aeruginosa*, *Water*  
424 *Research* 44 (2010) 1934-1940.

425 [12] Y.S. Lui, J.W. Qiu, Y.L. Zhang, M.H. Wong, Y. Liang, Algal-derived organic  
426 matter as precursors of disinfection by-products and mutagens upon chlorination,  
427 *Water Research* 45 (2011) 1454-1462.

428 [13] X. Yang, W. Guo, Q. Shen, Formation of disinfection byproducts from  
429 chlor(am)ination of algal organic matter, *Journal of hazardous materials* 197 (2011)  
430 378-388.

- 431 [14] J. Edzwald, Algae, bubbles, coagulants, and dissolved air flotation, *Water*  
432 *Science & Technology* 27 (1993) 67-81.
- 433 [15] C.W. Chow, M. Drikas, J. House, M.D. Burch, R.M. Velzeboer, The impact of  
434 conventional water treatment processes on cells of the cyanobacterium *Microcystis*  
435 *aeruginosa*, *Water Research* 33 (1999) 3253-3262.
- 436 [16] M.R. Teixeira, M.J. Rosa, Comparing dissolved air flotation and conventional  
437 sedimentation to remove cyanobacterial cells of *Microcystis aeruginosa*, *Separation*  
438 *and Purification Technology* 52 (2006) 84-94.
- 439 [17] Q. Shen, J. Zhu, L. Cheng, J. Zhang, Z. Zhang, X. Xu, Enhanced algae removal  
440 by drinking water treatment of chlorination coupled with coagulation, *Desalination*  
441 271 (2011) 236-240.
- 442 [18] D.L. Widrig, K.A. Gray, K.S. McAuliffe, Removal of algal-derived organic  
443 material by preozonation and coagulation: Monitoring changes in organic quality by  
444 pyrolysis-GC-MS, *Water Research* 30 (1996) 2621- 2632.
- 445 [19] M. Pivokonsky, O. Kloucek, L. Pivokonska, Evaluation of the production,  
446 composition and aluminum and iron complexation of algogenic organic matter, *Water*  
447 *Research* 40 (2006) 3045-3052.
- 448 [20] M. Ma, R. Liu, H. Liu, J. Qu, W. Jefferson, Effects and mechanisms of  
449 pre-chlorination on *Microcystis aeruginosa* removal by alum coagulation:  
450 Significance of the released intracellular organic matter, *Separation and Purification*  
451 *Technology* 86 (2012) 19-25.

452 [21] H. Hori, A. Yamamoto, E. Hayakawa, S. Taniyasu, N. Yamashita, S. Kutsuna, H.  
453 kiatagawa, R. Arakawa, Efficient decomposition of environmentally persistent  
454 perfluorocarboxylic acids by use of persulfate as a photochemical oxidant,  
455 Environmental Science & Technology 39 (2005) 2383-2388.

456 [22] A. Tsitonaki, B.F. Smets, P.L. Bjerg, Effects of heat-activated persulfate oxidation  
457 on soil microorganisms, Water Research 42 (2008) 1013-1022.

458 [23] H. Liu, T.A. Bruton, F.M. Doyle, D.L. Sedlak, In Situ Chemical Oxidation of  
459 Contaminated Groundwater by Persulfate: Decomposition by Fe(III)- and  
460 Mn(IV)-Containing Oxides and Aquifer Materials, Environmental Science &  
461 Technology 48 (2014) 10330-10336.

462 [24] S. Su, W. Guo, C. Yi, Y. Leng, Z. Ma, Degradation of amoxicillin in aqueous  
463 solution using sulphate radicals under ultrasound irradiation, Ultrason Sonochem 19  
464 (2012) 469-474.

465 [25] Y. Ji, Y. Fan, K. Liu, D. Kong, J. Lu, Thermo activated persulfate oxidation of  
466 antibiotic sulfamethoxazole and structurally related compounds, Water Research 87  
467 (2015) 1-9.

468 [26] T.K. Lau, W. Chu, N.J. Graham, The aqueous degradation of butylated  
469 hydroxyanisole by UV/S<sub>2</sub>O<sub>8</sub><sup>2-</sup>: study of reaction mechanisms via dimerization and  
470 mineralization, Environmental Science & Technology 41 (2007) 613-619.

471 [27] P. Neta, V. Madhavan, H. Zemel, R.W. Fessenden, Rate constants and mechanism  
472 of reaction of sulfate radical anion with aromatic compounds, Journal of the American



473 Chemical Society 99 (1977) 163-164.

474 [28] S. Ahn, T.D. Peterson, J. Righter, D.M. Miles, P.G. Tratnyek, Disinfection of  
475 ballast water with iron activated persulfate, *Environmental Science & Technology* 47  
476 (2013) 11717-11725.

477 [29] G. Mark, M.N. Schuchmann, H.P. Schuchmann, C. von Sonntag, The photolysis  
478 of potassium peroxodisulphate in aqueous solution in the presence of tert-butanol: a  
479 simple actinometer for 254 nm radiation, *Journal of Photochemistry and Photobiology*  
480 *A: Chemistry* 55 (1990) 157-168.

481 [30] P. Neta, R.E. Huie, A.B. Ross, Rate constants for reactions of inorganic radicals  
482 in aqueous solution, *Journal of Physical and Chemical Reference Data* 17 (1988)  
483 1027-1284.

484 [31] Y. Tao, X. Zhang, D.W. Au, X. Mao, K. Yuan, The effects of sub-lethal UV-C  
485 irradiation on growth and cell integrity of cyanobacteria and green algae,  
486 *Chemosphere* 78 (2010) 541-547.

487 [32] H. Ou, N. Gao, Y. Deng, H. Wang, H. Zhang, Inactivation and degradation of  
488 *Microcystis aeruginosa* by UV-C irradiation, *Chemosphere* 85 (2011) 1192-1198.

489 [33] M.G. Antoniou, A. Armah, D.D. Dionysiou, Degradation of microcystin-LR  
490 using sulfate radicals generated through photolysis, thermolysis and e<sup>-</sup> transfer  
491 mechanisms, *Applied Catalysis B: Environmental* 96 (2010) 290-298.

492 [34] N.S. Shah, X. He, H.M. Khan, J.A. Khan, K.E. O'Shea, D.L. Boccelli, D.D.  
493 Dionysiou, Efficient removal of endosulfan from aqueous solution by UV-C/peroxides:

494 a comparative study, *Journal of hazardous materials* 263 (2013) 584-592.

495 [35] V.K. Sharma, T.M. Triantis, M.G. Antoniou, X. He, M. Pelaez, C. Han, W. Song,  
496 K.E. O'Shea, A. Armah, T. Kaloudis, Destruction of microcystins by conventional and  
497 advanced oxidation processes: a review, *Separation and Purification Technology* 91  
498 (2012) 3-17.

499 [36] P. Xie, J. Ma, J. Fang, Y. Guan, S. Yue, X. Li, L. Chen, Comparison of  
500 Permanganate Preoxidation and Preozonation on Algae Containing Water: Cell  
501 Integrity, Characteristics, and Chlorinated Disinfection Byproduct Formation,  
502 *Environmental Science & Technology* 47 (2013) 14051-14061.

503 [37] R. Rippka, J. Deruelles, J.B. Waterbury, M. Herdman, R.Y. Stanier, Generic  
504 assignments, strain histories and properties of pure cultures of cyanobacteria, *Journal*  
505 *of General microbiology* 111 (1979) 1-61.

506 [38] P. Li, Y. Song, S. Yu, Removal of *Microcystis aeruginosa* using hydrodynamic  
507 cavitation: Performance and mechanisms, *Water Research* 62 (2014) 241-248.

508 [39] R.O. Rahn, Potassium iodide as a chemical actinometer for 254 nm radiation: use  
509 of iodate as an electron scavenger, *Photochemistry and Photobiology* 66 (1997)  
510 450-455.

511 [40] W. Liang, J. Qu, L. Chen, H. Liu, P. Lei, Inactivation of *Microcystis aeruginosa*  
512 by continuous electrochemical cycling process in tube using Ti/RuO<sub>2</sub> electrodes,  
513 *Environmental Science & Technology* 39 (2005) 4633-4639.

514 [41] G. Zhang, P. Zhang, B. Wang, H. Liu, Ultrasonic frequency effects on the

515 removal of *Microcystis aeruginosa*, *Ultrasonics Sonochemistry* 13 (2006) 446-450.

516 [42] O. Holm-Hansen, B. Riemann, Chlorophyll a determination: improvements in  
517 methodology, *Oikos* (1978) 438-447.

518 [43] R.I. Daly, L. Ho, J.D. Brookes, Effect of chlorination on *Microcystis aeruginosa*  
519 cell integrity and subsequent microcystin release and degradation, *Environmental*  
520 *Science & Technology* 41 (2007) 4447-4453.

521 [44] I. Kolthoff, E. Carr, Volumetric determination of persulfate in presence of organic  
522 substances, *Analytical Chemistry* 25 (1953) 298-301.

523 [45] H. Ou, N. Gao, Y. Deng, J. Qiao, H. Wang, Immediate and long-term impacts of  
524 UV-C irradiation on photosynthetic capacity, survival and microcystin-LR release risk  
525 of *Microcystis aeruginosa*, *Water Research* 46 (2012) 1241-1250.

526 [46] P. Xie, J. Ma, W. Liu, J. Zou, S. Yue, X. Li, M.R. Wiesner, J. Fang, Removal of  
527 2-MIB and geosmin using UV/persulfate: Contributions of hydroxyl and sulfate  
528 radicals, *Water research* 69 (2015) 223-233.

529 [47] X.-Y. Yu, Z.-C. Bao, J.R. Barker, Free Radical Reactions Involving Cl, Cl<sub>2</sub><sup>-</sup>, and  
530 SO<sub>4</sub><sup>-</sup> in the 248 nm Photolysis of Aqueous Solutions Containing S<sub>2</sub>O<sub>8</sub><sup>2-</sup> and Cl, *The*  
531 *Journal of Physical Chemistry A* 108 (2004) 295-308.

532 [48] G.P. Anipsitakis, D.D. Dionysiou, M.A. Gonzalez, Cobalt-mediated activation of  
533 peroxymonosulfate and sulfate radical attack on phenolic compounds. Implications of  
534 chloride ions, *Environmental science & technology* 40 (2006) 1000-1007.

535 [49] H.V. Lutze, N. Kerlin, T.C. Schmidt, Sulfate radical-based water treatment in

536 presence of chloride: Formation of chlorate, inter-conversion of sulfate radicals into  
537 hydroxyl radicals and influence of bicarbonate, *Water research* 72 (2015) 349-360.

538 [50] Y. Tao, X. Mao, J. Hu, H.O. Mok, L. Wang, D.W. Au, J. Zhu, X. Zhang,  
539 Mechanisms of photosynthetic inactivation on growth suppression of *Microcystis*  
540 *aeruginosa* under UV-C stress, *Chemosphere* 93 (2013) 637-644.

541 [51] H. Liu, T.A. Bruton, W. Li, J. Van Buren, C. Prasse, F.M. Doyle, D.L. Sedlak,  
542 Oxidation of Benzene by Persulfate in the Presence of Fe (III)-and Mn  
543 (IV)-Containing Oxides: Stoichiometric Efficiency and Transformation Products,  
544 *Environmental science & technology* (2015).

545 [52] R. Szabó, C. Megyeri, E. Illés, K. Gajda-Schranz, P. Mazellier, A. Dombi,  
546 Phototransformation of ibuprofen and ketoprofen in aqueous solutions, *Chemosphere*  
547 84 (2011) 1658-1663.

548 [53] B. Lee, M. Lee, Decomposition of 2, 4, 6-trinitrotoluene (TNT) by gamma  
549 irradiation, *Environmental science & technology* 39 (2005) 9278-9285.

550 [54] D. Guimaraes, V.A. Leao, Batch and fixed-bed assessment of sulphate removal  
551 by the weak base ion exchange resin Amberlyst A21, *Journal of hazardous materials*  
552 280 (2014) 209-215.

553 [55] M. Melkonian, Release of Dissolved Organic Substances by the Green Alga  
554 *Fritschiella tuberosa* Iyeng. (Chaetophorineae) During Different Growth Phases,  
555 *Zeitschrift für Pflanzenphysiologie* 94 (1979) 125-133.

556 [56] J. Huang, N. Graham, M.R. Templeton, Y. Zhang, C. Collins, M. Nieuwenhuijsen,

557 A comparison of the role of two blue-green algae in THM and HAA formation, *Water*  
558 *Research* 43 (2009) 3009-3018.

559 [57] H. Xu, H. Jiang, UV-induced photochemical heterogeneity of dissolved and  
560 attached organic matter associated with cyanobacterial blooms in a eutrophic  
561 freshwater lake, *Water research* 47 (2013) 6506-6515.

562 [58] H. Sakai, H. Katayama, K. Oguma, S. Ohgaki, Kinetics of *Microcystis*  
563 *aeruginosa* Growth and Intracellular Microcystins Release after UV Irradiation,  
564 *Environmental Science & Technology* 43 (2009) 896-901.

565 [59] N. Hudson, A. Baker, D. Reynolds, Fluorescence analysis of dissolved organic  
566 matter in natural, waste and polluted waters - A review, *River Research and*  
567 *Applications* 23 (2007) 631-649.

568 [60] P.G. Coble, Characterization of marine and terrestrial DOM in seawater using  
569 excitation. Emission matrix spectroscopy, *Marine Chemistry* 51 (1996) 325-346.

570 [61] W. Huang, H. Chu, B. Dong, J. Liu, Evaluation of different algogenic organic  
571 matters on the fouling of microfiltration membranes, *Desalination* 344 (2014)  
572 329-338.

573 [62] H. Ou, N. Gao, Y. Deng, J. Qiao, K. Zhang, T. Li, L. Dong, Mechanistic studies  
574 of *Microcystis aeruginosa* inactivation and degradation by UV-C irradiation and  
575 chlorination with poly-synchronous analyses, *Desalination* 272 (2011) 107-119.

576 [63] E.J. Rochelle-Newall, T.R. Fisher, Production of chromophoric dissolved organic  
577 matter fluorescence in marine and estuarine environments: an investigation into the

578 role of phytoplankton, *Marine Chemistry* 77 (2002) 7-21.

579 [64] L. Li, C. Shao, T.-F. Lin, J. Shen, S. Yu, R. Shang, D. Yin, K. Zhang, N. Gao,  
580 Kinetics of cell inactivation, toxin release, and degradation during permanganation of  
581 *Microcystis aeruginosa*, *Environmental science & technology* 48 (2014) 2885-2892.

582 [65] H. Miao, W. Tao, The mechanisms of ozonation on cyanobacteria and its toxins  
583 removal, *Separation and Purification Technology* 66 (2009) 187-193.

584 [66] R.P. Sinha, D.-P. Häder, UV-induced DNA damage and repair: a review,  
585 *Photochemical & Photobiological Sciences* 1 (2002) 225-236.

586 [67] J. Cadet, C. Anselmino, T. Douki, L. Voituriez, New trends in photobiology:  
587 Photochemistry of nucleic acids in cells, *Journal of Photochemistry and Photobiology*  
588 B: Biology 15 (1992) 277-298.